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5.15 Estuaries and Brackish Waters

5.15.1 Introduction

Estuaries can be defined as tidally influenced transition zones between marine and riverine environments. According to the Leeuwarden Declaration, the estuaries in the Trilateral Cooperation Area are delimited on the landward side by the mean brackish water limit and on the seaward side by the average 10 PSU isohaline at high water in the winter situation. However, in the present review the tidally influenced freshwater reaches are included (according to the definition of Fairbridge, 1980), as they form an integral part of the estuarine habitat (Schuchardt *et al.*, 1993a). Laterally, all areas up to the main dike or, where absent, the spring high tide water line, are included.

According to the above definition, there are

five estuaries in the Wadden Sea region: the Varde Å estuary in Denmark and the Eider, Elbe, Weser and Ems estuaries in Germany.

These estuaries are, on the one hand, of high relevance for the Wadden Sea ecosystem (input of nutrients and toxic substances, sediment dynamics, nursery and feeding area). On the other hand, the estuaries themselves are a specific habitat, characterized by strong variability and dynamics of key-factors such as salinity, tidal range, turbidity and others. From an ecological point of view, they are important, e.g. for the migration of a number of species and, additionally, they are inhabited by various brackish-water and, at least, estuary-endemic species, thus being of special importance for conservation purposes. However, in contrast to the Wadden Sea, the estuaries are strongly altered by human activities and only scat-

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Figure 5.19. Estuaries in the Wadden Sea region.

tered areas are protected as nature reserves.

The data describing the ecological situation is heterogenous and a positive correlation between length of the estuary and the amount of available data exists. Systematic long-term monitoring data only exist for water quality and, partly, macrozoobenthos. For biological data, only scattered information is available, resulting mainly from research programs of universities/authorities and, very recently, data sets obtained in the framework of environmental impact assessments.

5.15.2 Brief characterization

The estuaries of the Varde Å, Eider, Elbe, Weser and Ems are mesotidal coastal plain estuaries debouching into the Wadden Sea (North Sea) (Figure 5.19).

have strongly affected virtually all estuaries under consideration except the Varde Å. The main reasons were coastal protection and land reclamation and adaptation to increasing ship numbers and sizes.

Deepening of the shipping routes

Since the beginning of the century, the estuaries of Elbe and Weser have been successively adjusted to the increase in the average size of the world merchant fleet to allow ocean liners to reach Hamburg and Bremen. In the Ems estuary, the driving forces are the port of Emden and the transport of newly built ships from the Papenburg shipyard to Emden. The adaptation of the estuaries of Elbe, Weser and Ems was accomplished by the excavation of a straight navigable channel to concen-

Table 5.9. Compilation of main features, describing the estuaries of Varde Å, Eider, Elbe, Weser and Ems. Data from Schuchardt, 1995; Pedersen, pers. comm.

	Varde Å	Eider	Elbe	Weser	Ems
drainage area upstream of tidal weir (km ² ·10 ³)	1.08	2	135	38	13
mean river discharge (m ³ /s)	13	23	725	323	125
mean tidal range at tidal weir (m)	1.3	2	2.4	4.1	2.8
length of the inner estuary (km)(definition see text)	17	21	120	70	50

Morphologically, the river mouth can be divided into two sections: a river-like inner part (including the tidal freshwater, the oligohaline and parts of the mesohaline reaches) between a tidal weir and the outer part. This outer part is characterized by a funnel-like morphology with very extended tidal flats, being part of the Wadden Sea. From this general structure, the Ems estuary differs due to the Dollart, a brackish bay, the Eider estuary, due to the construction of a storm-surge-barrier and the Varde Å, where mixing of river and sea water takes place normally in the Ho Bugt seaward of the narrow river (Duinker *et al.*, 1980). According to the definition given above, this paper deals mainly with the inner parts of the estuaries.

The estuaries vary considerably in size, length and river discharge (Table 5.9). Except for the Varde Å (Olesen, 1993) the adjacent lowlands in all estuaries are protected by dikes against flooding, either from the sea or from the river, and they are of great economic importance for shipping, agriculture and industrial purposes and have been altered by human activities.

5.15.3 Morphological modifications

Morphological alterations of rivers and estuaries

trate and accelerate the water masses. Concentration was increased by groynes and by the cutting and filling in of sidearms and ramifications. Shallows and gravel banks were often removed or elevated and connected to each other. The shipping channel was smoothed; roughness and energy-dissipation reduced and given a funnel-like shape, the narrow end leading into the upstream laying harbors (Schirmer, 1996; Kausch, 1996).

The changes are most extreme in the Weser estuary, especially when compared to the natural width of the estuary and the freshwater discharge. The Ems is similarly impacted; the Dollard much less than the Elbe estuary, where the shaping of the channel left much of the estuarine topography unchanged.

The stepwise deepening and canalization of the outer and inner estuaries of Elbe, Weser and Ems led to a marked increase in tidal amplitude. The correlation between tidal amplitude within an estuary and its morphology is fairly high and allows the use of this well-documented parameter as an indicator of the anthropogenic impact on the estuaries considered here. Another advantage of monitoring and evaluating tidal amplitudes lies in its importance as an ecologically relevant factor.

Figure 5.20 gives an impression of the changes in tidal range between 1880 and 1995. About 75%

of the increase in tidal amplitude results from a drop of the mean low water level, about 25% from a rise of mean high water (Harten, 1979). The increase in the Weser is extreme due to the excessive canalization of the estuary and a reflection of the tidal wave at the weir in Bremen-Hemelingen.

The increased current velocities transport increasing amounts of suspended solids into the side waters. This material is accreted there and causes a slow but steady rise of the bottom level. This happens below the low water line as well as on tidal flats, thus transforming shallow water areas into tidal flats.

The strong tidal currents, together with wave action from passing ships, tend to erode the banks and, thus, enforce bank protection measures, especially in the narrower estuaries of Weser and Ems. In the Weser estuary, e.g. about 60 % of the banks along the main channel between Bremen and Bremerhaven are protected, mainly by rock material (Schuchardt *et al.*, 1984).

Maintenance dredging is necessary in all of the canalized estuaries creating environmental impact during dredging and during dumping. It is least within the shipping channel of the Weser estuary compared to the Elbe and Ems as indicated in Table 5.10. The removed sediments from the channels of Elbe, Weser and Ems are either transported into the outer estuaries and dumped next to the shipping channel or dumped in the inner estuary

in stretches which are over-deepened due to erosion. With respect to the material balance between estuaries and the Wadden Sea, the removal of sand (and silt) from the estuaries raises critical questions on the long-term consequences for the Wadden Sea sediment budget (Höpner, 1994).

Coastal protection and land reclamation

Another main factor for far-reaching modifications in all estuaries is coastal protection. Linear dikes along the estuarine banks have been built since the 11th and 12th century (Meier, 1996). Until the 19th century, Bremen and parts of Hamburg, however, were prone to flooding by high discharges of Weser and Elbe. Only after the dredging, did the storm surges from the North Sea become the greatest danger and dikes had to be reinforced. Many of these dikes still follow the old lines but land reclamation and shortening of dike length was a permanent issue: the Eider estuary lost about 12 km² of forelands in 1972 (Katinger Watt), the Elbe estuary lost about 140 km² between 1896 and 1982, the Weser lost about 32 km² between 1880 and 1955 and the Ems/Dollard-region has lost about 21 km² since 1922 (Schuchardt *et al.*, 1993b). Major land reclamation after 1975 is only found along the Elbe estuary (Hasedorfer Marsch 20 km², Asseler- und Krautsand 26 km² (ARGE Elbe, 1984).

The importance of dikes along the estuaries has been steadily growing because of

Figure 5.20. Changes in tidal range between 1880 and 1995 in the Eider (gauges Tönning and Friedrichstadt), Elbe (gauge Hamburg St. Pauli), Weser (gauge Bremen Oslebshausen) and Ems (gauge Herbrum) (5-year-running mean). According to Schuchardt, 1995.

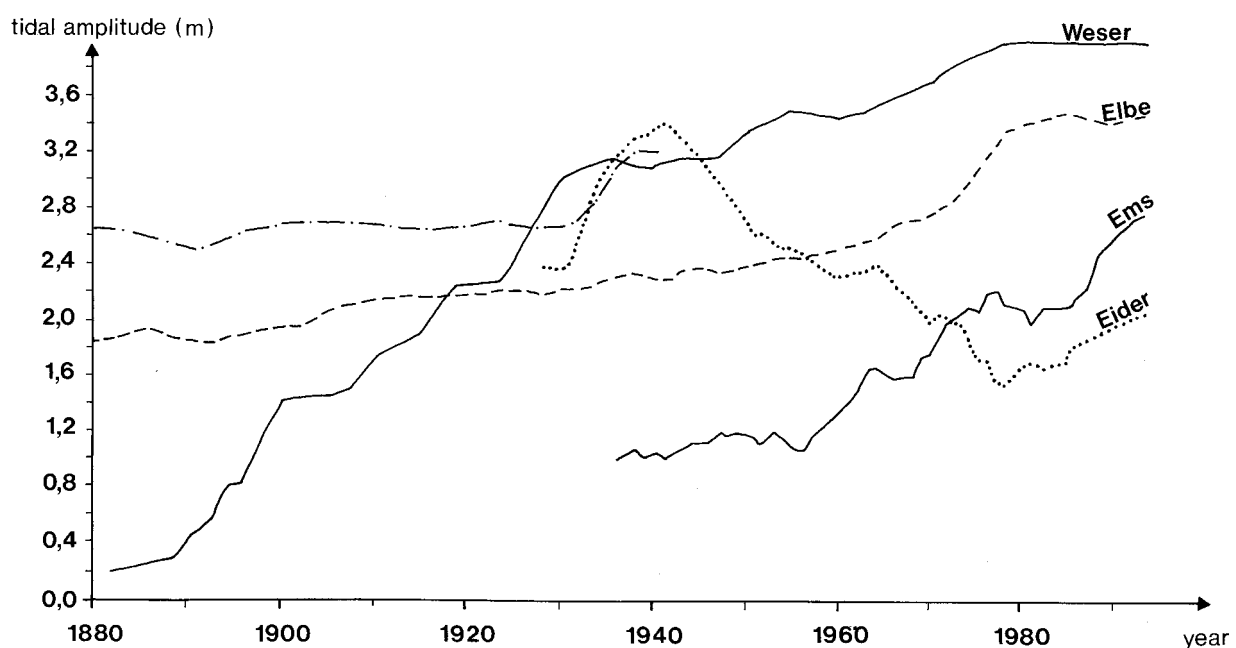


Table 5.10. Mean annual volumes ($10^6\text{m}^3/\text{yr}$) of maintenance dredging in the shipping channels of Elbe, Weser and Ems Estuaries between 1981 and 1992 (after Farke, 1994) and sediment accumulation in the harbors (according to Nasner, 1992).

	Elbe	Weser	Ems
av. volume in channel (min-max)	12.8 (9.5-16.5)	3.4 (1.3-7.9)	7.3 (0-10.5)
av. sedimentations in harbour area	2.4 (Cuxhaven+Brunsbüttel+Hamburg)	1.0 (Bremerhaven+Bremen)	0.6 (Emden)

- i) the secular sea level rise;
- ii) the enhancement of storm surges by dredging and channeling;
- iii) the construction of storm surge barriers on the tributaries which reduced the flood basin dramatically and raised the flood level, e.g. in the Weser more than 0.5 m;
- iv.) strong subsidence of the marshes due to soil mineralization and drainage.

Additional losses of interaction between river and marsh were caused by low summer dikes constructed to protect forelands in agricultural use. As a result of diking and channeling, the hydrological and sedimentary interaction between the estuaries and their marshes has been interrupted to a maximum degree. The hydrological interaction is highly controlled. Drainage and irrigation are accomplished by sluices and pumps, biological interaction is restricted (Garms, 1961).

The situation is somewhat different for the Eider estuary. In addition to diking, the morphology has changed strongly due to a reduction of the catchment area by the Kiel Canal by about 30% (1895) and the construction of a tidal barrage at Nordfeld (1936). The cross-section below Nordfeld shrunk to 5 - 10% of the initial width due to siltation. The construction of a storm-surge barrier at the mouth of the Eider (Vollerwiek) in 1973 and the diking of the Katinger Watt reduced the cross-section even more. Today, the incoming tidal flood is actively reduced to prevent further siltation (Fock and Ricklefs, 1996).

5.15.4 Estuarine habitats: changes in quantity and quality

The morphological and physical alterations of the estuaries had an immense impact on habitat structure and quality both in the marshes, the amphibious zone and in the aquatic environment.

The Forelands

Foreland areas between the mean spring tide high water line and the winter dikes are frequently flooded and are of special ecological importance. About two to three quarters of the forelands along the inner estuaries of Elbe, Weser and Ems were

lost during the last century (Schuchardt *et al.*, 1993b). Many of the remaining marshes are protected by low summer dikes, which prevent much of the everyday interaction between river and marsh. This is being amplified by the lowering of the ground water table due to the lowering of the tidal low water and midwater levels in the estuaries of Elbe, Weser and Ems and by active drainage for agricultural use.

Tidal flats and shallow waters

The increase in tidal amplitude due to the deepening of the estuaries could have created more tidal flats and shallow water areas (with fewer than 2 m depth below tidal low water). However, the deposition of dredged sediments within the tidal area, the reclamation of land, the widening of the deep shipping channel and the protection of banks have altogether strongly reduced these areas. The losses are worst within the Weser estuary between Bremen and Bremerhaven and less pronounced in the Elbe (Schirmer, 1994).

5.15.5 Ecological structure: abiotic parameters

The estuarine habitat can be characterized as a very dynamic environment, governed by steep gradients of salinity and turbidity, the tidal rise and fall and resulting currents, long water residence time compared to the river, partly pronounced oxygen deficits, relatively high nutrient concentrations and, before diking, extended lowlands flooded both due to riverine peak discharge and during storm flood situations.

Salinity gradient

The freshwater-seawater interface is characterized by the mixing of sea and river water resulting in a steep, strongly variable salinity gradient. The position of the resulting brackish habitat varies with the tides and due to changes in riverine discharge. The gradient has been strongly altered in the Eider estuary due to a tidal weir and a storm surge barrier with some consequences for the brackish water community (Fock and Heydemann, 1995). In the Weser estuary, an upstream shift of

the brackish water has been found (Grabemann *et al.*, 1983). Morphological changes due to canalization could not be clearly identified as main reasons, but must be assumed. This is similar to the situation in the Elbe estuary (Bergemann, 1995).

Turbidity zone

A zone of high turbidity associated with the upstream limit of the salinity gradient is typical for all estuaries under consideration (Wellershaus, 1981; Grabemann *et al.*, 1996). The downstream border of this reach may be set at a salinity of about 8 PSU. In this turbidity zone, phytoplankton production is restricted due to light limitation (Schuchardt and Schirmer, 1991). A relation between extensive dredging and dumping activities or estuarine canalization and increased turbidity has been discussed (Wellershaus, 1986).

Temperature

Water temperature as an ecological master factor is locally altered due to the input of cooling water (Müller *et al.*, 1990). Several large power plants are located along the inner estuaries of Elbe and Weser.

Tides

The tides are an essential feature of an estuary (aestus = tide). However, in the Wadden Sea estuaries, the tidal amplitude has been strongly changed due to canalization and other activities (Schuchardt, 1995). Especially in the tidal freshwater reaches, the habitat structure has been altered (Schirmer, 1994). The decrease of the mean low tide level has led to a nearly complete transformation of shallow areas to tidal flats in the inner Weser estuary over the last 100 years. In the adjacent lowlands, tidal influence has been reduced due to diking and water management (Heckman, 1984). Thus, the function of these areas for the estuary, like retention of suspended matter and nutrients, has been diminished (Heckman and Kausch, 1996).

Residence time

The water residence time within the estuary is an important feature preventing downstream loss of planktonic species and increasing oxygen demand due to C- and N-mineralization. It depends mainly on the relation between the river discharge and the estuarine water volume. The water volume has been increased in the estuaries of Elbe, Weser and Ems due to deepening and decreased in the Eider estuary as a result of siltation after construction of the tidal weir (Fock and Heydemann, 1995).

Oxygen

Slight oxygen deficits have been observed to be a typical estuarine feature associated with the accumulation of suspended particulate matter in the estuarine turbidity zone (Martin and Brun-Cottan, 1988). In the Eider estuary, oxygen deficits are only minor, indeed. However, in the Weser and Elbe estuaries, it had been strongly enhanced as a result of waste water input over the last decades. In the Elbe estuary, recently, a reduction of the summer oxygen deficits can be observed (Reincke, 1995). This is both a result of improving waste water treatment in Hamburg and a reduction of oxygen-consuming substances from upstream. Also, in the Weser estuary, oxygen deficits have been reduced since waste water treatment in Bremen has been improved (Schuchardt *et al.*, 1989). A significant deterioration has been observed in the Ems estuary in the last years probably due to excessive dredging activities, leading to an impoverishment of the estuarine biocoenosis (IBL, 1997).

Nutrients and pollutants

The nutrient concentrations in the river-dominated estuaries are clearly elevated compared to a pre-industrial situation, which has been estimated for the Elbe estuary (Flügge *et al.*, 1989). However, there exist some differences between the estuaries under consideration. This is detailed in Chapter 4.

For concentrations of heavy metals and trace organic compounds see Chapter 4.

5.15.6 Ecological structure: biota

The estuarine biocoenosis is substantially governed by physical factors such as tides (tidal range and currents), water depth and river discharge as well as by the salinity gradient. Suspended particulate matter dynamics, morphodynamics and sediment properties are also important. However, human uses have altered the ecological structure of all estuaries (Schuchardt *et al.*, 1993b, Höpner, 1995).

Phytoplankton

Phytoplankton biomass shows a steep gradient within the estuaries with a peak in the tidal freshwater reach and minimum values in the turbidity zone. Light limitation has been found to be of greater importance than salinity or nutrients. Biomass, upstream of the turbidity cloud, is mainly formed by the estuarine diatom *Actinocyclus normanii* during summer and autumn, whereas riverine species (*Asterionella*, *Synedra*) are domi-

nating during spring (Schulz, 1961, Schuchardt and Schirmer, 1991).

Biomass in shallow side-arms has been found to be several-fold higher than in the deep main channel (ARGE Elbe, 1984), indicating the consequences of subsequent reduction of shallow areas due to deepening of the estuaries.

For the Weser estuary, some shifts of species composition in the last 50 years have been documented, mainly as a result of increasing salt-concentration from potassium-mining and due to nutrient increase (Haesloop and Schuchardt, 1995). However, recently, a shift in species composition has been observed in the Weser River as a result of a reduction of the anthropogenic salt burden (Bäthe and Preuß, 1994).

Microphytobenthos

Microphytobenthos in the eulittoral has been analyzed by Gätje (1992) for the Elbe estuary. It is dominated by widely distributed diatom species. Interactions with the phytoplankton via sedimentation and resuspension have been found to be of some importance (Kies *et al.*, 1992). Reduction of tidal areas due to the deepening for sea-going vessels and other use have reduced their habitat (Kies *et al.*, 1992). Of some importance are benthic (macro)-algae, mostly species of the genera *Vaucheria*, *Enteromorpha* and *Blidingia*.

Vegetation

Submersed macrophytes are nearly completely absent in all of the estuaries. In the oligohaline and mesohaline reaches, this is due to natural tidal amplitude, salinity and turbidity. In the limnic reaches, this is mainly a result of anthropogenic increase of the tidal range (Runge, 1981 for the Ems; Cordes, 1993 for the Weser). During mappings at the beginning of the century, species such as *Potamogeton pectinatus*, *P. lucens* and *Myriophyllum spicatum* were abundant.

As a result of diking, agricultural use of foreland and disposal of dredged material on land, small bodies of waters above the high tide water line have been reduced. In such waters, submersed macrophytes might be still abundant due to the absence of tidal rise and fall.

Eulittoral vegetation (mainly reeds) is still of importance in the estuaries (Kies *et al.*, 1992). Although strongly reduced, compared with the situation at the beginning of the century (e.g. Preisinger, 1991, Claus *et al.*, 1994), the typical vertical and longitudinal zoning of *Schoenoplectus tabernaemontani*, *Bolboschoenus maritimus* and *Phragmites australis* is still present (e.g. Kötter, 1961). In the Eider estuary, steepening and sea-

ward displacement of the freshwater-seawater interface, resulting from several constructions (stormsurge-barrier, tidal weir), has led to strong sedimentation and subsequent downstream increase of reeds in the area of former extended salt marshes (Fock and Heydemann, 1995).

Foreland vegetation above MHW is characterized by agricultural use and the loss of formerly widely distributed vegetation communities and species (Cordes, 1993; Claus *et al.*, 1994). Endemic species such as *Oenanthe conioides* and *Dechampsia wibeliana* adapted to tidal influence have become rare (Heckmann and Kausch, 1996). Riparian forests, widely distributed on the riverbanks of the tidal freshwater reaches before the settling of man (Behre, 1985) are nearly completely lacking today.

Zooplankton

Zooplankton communities in the estuaries under consideration have been found to be similar (Kühl and Mann, 1981). The inner and middle estuarine reaches are dominated by the calanoid copepod *Eurytemora affinis* (Kausch and Peitsch, 1992; Haesloop and Schuchardt, 1995). During summer, the rotifers *Keratella* spp. and *Brachionus* spp. and some cladocera may be of some quantitative importance, especially in the innermost estuary. In the turbidity zone, the rotifers decrease strongly and biomass is dominated by *Eurytemora* during the year. Due to the invasion of some marine species (e.g. *Synchaeta*), there is no minimum in the species number found. In the water column, several mobile crustacea like *Crangon crangon*, *Palaemon longirostris*, *Gammarus* spp. and *Neomysis integer* have been found to be of quantitative importance (Fiedler, 1991).

Similar to the distribution of phytoplankton, abundance of zooplankton in areas with reduced current velocities like harbors open to the tide, can be several times higher (Schuchardt *et al.*, 1995).

Significant long-term changes in species composition have not been found for the inner Weser estuary (Haesloop and Schuchardt, 1995). However, reduction of shallow sidearms due to canalization caused strong habitat reduction.

Macrozoobenthos

Macrozoobenthos is relatively well documented for the four estuaries. In general, high densities and low diversity are typical for the brackish zone, whereas lower densities and increasing diversity are typical for the outer estuary. In the tidal freshwater reaches upstream, diversity is low. This is mainly due to the tides, as can be shown by comparing the situation upstream of the tidal weir

(Scholle and Schuchardt, 1997). However, motile benthic species (mostly crustacea) are of some importance (Haesloop, 1990; Arntz *et al.*, 1992).

In the Eider estuary, the salinity gradient has been strongly altered (see above) resulting in a reduction of the length of the brackish reach and a steepening of the salinity gradient. Marine species such as *Nereis diversicolor*, *Arenicola marina* and *Cerastoderma edule* have retreated downstream (Fock and Heydemann, 1995) and the number of genuine brackish species is low (Michaelis *et al.*, 1992). However, recent surveys (Fock, 1996) have documented the occurrence of further brackish water species (*Tubificoides heterochaetus*, *Paranais frici*, *P. botniensis*). Quantitatively, the low salinity region is characterized by *Marenzelleria* spp. (*Viridis/wireni* complex), *Heteromastus filiformis* and *Macoma balthica*.

Although the former situation of the macrozoobenthos in the Elbe estuary is well documented, recent published data are scattered. Stony embankments have been sampled by Gaumert and Spieker (1995) showing higher densities and diversity in the low salinity region (dominated by *Balanus improvisus*) than in the tidal freshwater reach (dominated by *Dreissena polymorpha* and *Gammarus zaddachi*). According to Nehring and Leuchs (1996), the low salinity region is characterized by *Marenzelleria* spp., *Boccardiella redeki* and *Neomysis integer*. Riedel-Lorje *et al.* (1995) and Krieg (1996) have described some upstream extension of brackish water species such as *Bathyporeia pilosa* and *Corophium volutator* and concluded that there has been an upstream movement of the brackish water zone, perhaps as a result of engineering works in the estuary. However, this item is still open to scientific discussion.

In the Weser estuary, the situation is well documented by the results of, e.g. Haesloop (1990), Kolbe (1995b) and Gosselck and Prena (1996). Due to elevated conductivity (resulting from potassium mining upstream) in the tidal freshwater reach, the benthic fauna is dominated by several brackish and marine crustaceans, most of them invading in large numbers from the natural brackish region in spring and summer. Among them, *Gammarus zaddachi*, *Corophium lacustre*, *Neomysis integer*, *Palaemon longirostris* and *Crangon crangon* are of particular importance (Haesloop and Schuchardt, 1995) as well as *Balanus improvisus* and *Marenzelleria* spp. (Nehring and Leuchs, 1996). In the natural brackish reach, further species such as *Balanus improvisus*, *G. salinus*, brackish oligochaetes like *Peloscolex heterochaetus* and *Tubifex costatus* as well as some polychaetes are

of some importance as well as a number of marine species. Analyzing the existing data sets, a long-term data set compiled by Kolbe (1995b) is of special interest. Data on infauna from the natural brackish water zone near Bremerhaven over 20 years show that strong variability of species diversity and abundance is a typical feature of estuarine communities, making it very difficult to differentiate between natural dynamics and manmade changes (analyzing estuarine community data).

In the freshwater and the oligohaline reaches of the Ems estuary, species number, biomass and densities are lower than in the mesohaline reaches (IBL, 1994). Species most present in the tidal freshwater reach are *Limnodrilus hoffmeisteri* and other tubifids as well as *Gammarus zaddachi*. Species composition in the oligohaline reaches is similar, whereas in the mesohaline reach brackish species like *Marenzelleria* spp., *Tubifex costatus*, *Bathyporeia pilosa* and some marine species are important. However, there is some evidence that reduced habitat diversity and dredging have reduced the number of species (Dittmer, 1981; Rhode, 1982; IBL, 1994). Arntz *et al.* (1992) have described the quantitative importance of motile macrofauna species such as *Paramysis kervillei* and *Neomysis integer*. However, recently, a decline of diversity and abundance, probably due to dredging activities and anoxia, has become obvious (IBL, 1997).

Comparison of recent data with older ones from the beginning of the century, reveals a strong reduction of species number in the innermost Weser estuary, which is mainly the result of morphological changes, whereas in the middle reaches, changes are less significant (Haesloop and Schuchardt, 1995). It can be assumed that this is a general tendency in the estuaries under consideration (for the Elbe see BfG, 1994). However, due to the intrusion of a number of "new" species (neozoa), these losses have partly been compensated (see Nehring and Leuchs, 1999). These alien species seem to have become integrated into the food web without out-competing other species (in estuarine sediment dumping areas *Marenzelleria* spp. becomes one of the dominant forms; Leuchs *et al.*, 1996).

Fish

Fish is relatively well documented as well, due to its great commercial importance in historical times. This importance has been reduced strongly, both due to a decrease or extinction of some fish species of commercial interest and changes in the behavior of consumers (Schuchardt *et al.*, 1985; Möller, 1989). However, in general, the estuaries

still sustain large fish stocks. The open boundaries to the marine as well as to the fluvial reaches and the high productivity are the most important prerequisites.

For the Eider estuary, only scattered data are available. According to Dehus (1987), euryhaline species (see Elbe section) are dominant in the upper estuary. There is some evidence that one main reason for the extinction of sturgeon (*Acipenser sturio*) was the construction of the tidal weir in 1936, isolating the adults from their upstream spawning grounds (Spratte, 1992). Due to the weir, the hydrographic situation upstream, in this former tidally influenced stretch, has been changed: rheophilic species such as lampern (*Lampetra fluviatilis*) have been reduced, whereas lenitic cyprinids are still abundant (Dehus, 1987; Spratte, 1992). A strong decline of eel (*Anguilla anguilla*) catches has been explained as a result of the storm-surge barrier (Fock, 1998).

The situation in the Elbe estuary is very well documented. In the inner Elbe estuary (Tideelbe), 77 species have recently been recorded, of which 31 are freshwater species (6 of them are allochthonous), 11 can be classified as euryhaline (one of them allochthonous) and 34 of them are marine species (Thiel, 1995). Four species, formerly abundant, are extinct. These are *Vimba vimba*, *Silurus glanis*, *Alosa alosa* and *Acipenser sturio*. Some other species have been strongly reduced in numbers. Quantitatively, the euryhaline species smelt (*Osmerus eperlanus*), twaite shad (*Alosa fallax*), flounder (*Platichthys flesus*), stickleback (*Gasterosteus aculeatus*) and eel (*Anguilla anguilla*), the marine species herring (*Clupea harengus*) and the freshwater species ruffe (*Gymnocephalus cernua*) are the most important. Biomass in the shallow side-arms is about one third higher than in the main channel giving an impression of the consequences of the subsequent reduction of shallow areas in all of the estuaries (see above). Longitudinally, no differences in fish abundance have been found, whereas biomass is increasing seaward due to the appearance of marine species. For the Elbe estuary, some impact of oxygen deficits on the distribution of smelt and flounder has been found: they are avoiding areas with concentrations lower than 3 mg/l. Today, the oxygen deficits during summer are less, due to the improvement of water quality (Reincke, 1995). The concentrations of heavy metals and, especially chlorinated hydrocarbons in eels, are still elevated in the Elbe estuary, resulting mainly from the input via some tributaries of the middle reaches of the Elbe (Reincke, 1995). This prompted a ban on marketing Elbe eel for human consumption. How-

ever, a reduction of contamination in the future can be expected. The influence of the tidal weir at Geesthacht on fish is documented (Beckedorf and Schubert, 1995). Impact of the entrainment of cooling water is described by Köhler (1981).

According to the recent compilation by Schirmer (1991) and Scheffel and Schirmer (1997), the fish community in the Weser estuary seems to be very similar to that in the Elbe estuary. The same species are rare or extinct (in addition to those already mentioned for the Elbe, also *Coregonus oxyrhynchus* and *Lota lota*) and a similar spectrum of species is dominating (euryhaline species and some cyprinids). A characteristic species in the innermost estuary is the ide (*Leuciscus idus*). The main reasons for the extinction of species like *Acipenser sturio* and others are the same as for the Elbe estuary (Busch *et al.*, 1989).

For the Ems estuary, dominance of the same euryhaline species as in the Elbe and Weser is documented (Arntz *et al.*, 1992). However, occurrence of *Alosa alosa* is still possible and for *Coregonus oxyrhynchus* there exist records up to the seventies (Gaumert and Kämmereit, 1993). Recently, the fish fauna seems to have been declining strongly (IBL, 1997). Whether this is due to excessive dredging activities or anoxia during summer or other factors is still unclear.

For the Varde Å, data are not available. However, the restocking of *Coregonus oxyrhynchus* should be mentioned (see Section 5.10).

Summarizing the situation in the estuaries, it can be stated that they are still (and partly again) populated by a diverse fish community and that they are still of importance as a nursery ground for Wadden Sea species (e.g. flounder *Platichthys flesus*, see Kerstan, 1991). However, some alterations of the fish communities due to anthropogenic impact are obvious. Some species (especially diadromous species) have become extinct or rare, the main reasons being overfishing, construction of weirs, destruction of spawning grounds and pollution. Also, reduction of shallow areas seems to be an impact factor. Recently, some positive developments like increase of smelt in the Elbe estuary as a result of the improvement of water quality can be stated. However, the situation in the Ems estuary requires special attention. During recent years, there seems to be a sharp decrease of species diversity and abundance. Measures improving the ecological situation are necessary.

Special attention must be given to the strong decrease of eel catches in the North Sea and, especially, to the reduction of young eel invading the estuaries from the North Sea.

5.15.7 Ecological functions of the estuaries

The most important ecological functions of estuaries result both from their open boundaries to marine and limnic reaches and from their own specific properties.

Linkage and gateway

The linkage and gateway function of estuaries between marine and limnic environments is an essential feature in the life cycle of several fish and invertebrate species (e.g. *Salmo salar*, *Anguilla anguilla* as the most important). This function is diminished in all estuaries due to the construction of tidal weirs (e.g. Beckedorf and Schubert, 1995).

Filter

The filter function of estuaries for particulate and partly dissolved matter (due to the increase of water-residence-time, partial retardation of particulate matter in the turbidity cloud and of sediment on estuarine flats and marshes, degradation of organic material) is of importance for the input of substances to the Wadden Sea and the North Sea. However, this function has been strongly reduced due to manmade morphological changes (Kausch *et al.*, 1991; Heckman and Kausch, 1996).

Habitat

The importance of the habitat function of estuaries for the genuine brackish water species has increased after the nearly complete destruction of the former gradual transition area between marine and limnic reaches along the coast due to diking and water management (Wolff, 1991). However, due to physical (morphological and hydrological) alterations of the estuaries this function is restricted (Michaelis *et al.*, 1992).

Sediment supply

The sediment supply function of the estuaries seems to be of importance for the sediment bal-

ance in the Wadden Sea (e.g. Höpner, 1994). The input has been reduced due to the construction of dams, maintenance dredging with deposition on land and sand extraction. However, quantitative importance cannot be assumed yet.

The trophic dynamics of the estuaries under consideration is still only partly understood and the scientific discussion of the relative importance of organic detritus versus phytoplankton for estuarine consumers is open (see Day *et al.*, 1989). Little comparable data on the productivity of estuaries is available. For the primary production of plankton, microphytobenthos and macrophytes, some data from the Elbe estuary are shown in Table 5.11. The high biomass and production of macrophytes per m² is remarkable, whereas turnover (p/b-ratio) is highest for phytoplankton. Taking the entire areas available in the Elbe estuary into account, the contribution of phytoplankton is relatively small, whereas the contribution of macrophytes is highest. This means that the strong reduction of estuarine areas populated by macrophytes, as mentioned above, has had considerable negative consequences also for the estuarine productivity.

Phytoplankton production in the estuary has been influenced by the deepening for sea-going vessels. Deepening has changed the relation between the euphotic zone and non-euphotic zone. This can only partly be compensated by the heterotrophic activities of estuarine phytoplankton (Kies *et al.*, 1992). For planktonic primary production, shallow areas are, thus, of special importance as can be shown in the remaining side-arms in the Elbe estuary (Schuchardt and Schirmer, 1991). However, such shallow areas are nearly completely lacking in the canalized Weser estuary and have been reduced in the other estuaries.

5.15.8 Other brackish waters

Diking and land reclamation all along the coast of the Wadden Sea have eliminated the broad

	biomass	production (gC/m ²)	p/b-ratio (gC/gC. d)	production (tC/yr)
phytoplankton	4.1	0.35	0.09	2,933
microphytobenthos	12.8	0.66	0.05	10,000
macrophytes	686.4	9.42	0.01	30,072

Table 5.11. Comparison of biomass and primary production for phytoplankton, microphytobenthos and macrophytes from the Elbe estuary in September 1987 per m² (Kies *et al.*, 1992) and for the whole area of the Elbe estuary between km 650 and 715 (Kies and Neugebohrn, 1994).

natural transitional zone with creeks, swamps and marshes (Behre, 1985) where freshwater draining from land and seawater mixed and created a chain of brackish water habitats between the big open estuaries. Only a few habitats of this kind, like the mouth of the Godel creek on the island of Föhr (Figure 5.19)(see Stock *et al.*, 1996), have been left. The concentration of drainage systems into a few sluices, with a combination of gates and powerful pumps, created small, isolated brackish water environments with very unstable conditions. Little and/or irregular freshwater flow, depending on intermittent pumping, is the main reason for the very poor ecological quality of these biotopes as stated by Michaelis *et al.* (1992). They checked eight smaller outlets and found only half as many brackish-water species (≤ 7) as in the riverine estuaries (≤ 14) (see also Wolff, 1991). Most of these show hardly any salinity gradient but a sharp salinity break and have been termed fresh-salt-water changes by den Hartog (1964, 1967). About 25 of these are found in the Dutch Wadden Sea (de Boer and Wolff, 1996). Tidal creeks on the salt marshes of the Wadden Sea islands are considered to be another category of "mini-estuaries". Although only small amounts of freshwater run off through these creeks (at least 15 in the Dutch Wadden Sea), a brackish gradient in these creeks might be present. On two islands, dikes have been removed to restore estuarine gradients and further measures are planned.

The former estuary of the river IJssel in the Dutch Wadden Sea became a freshwater lake in 1932. The freshwater of the IJssel is now discharged through the sluices in the Afsluitdijk, creating estuarine conditions in the western Wadden Sea. In the Dutch Wadden Sea, there are about 11 such semi-open 'estuaries, forming real barriers for organisms, which use these transitions zones for migration between the fresh and the marine environment.

A different type of brackish water is found in several polders along the German coast. Whilst most of the enclosed waters turned into freshwaters, in four of them, brackish water prevails due to special water management. However, all suffer from eutrophication, algal blooms and oxygen deficits (Hagge, 1994; Nehring *et al.*, 1998).

Another type of brackish water is represented by the artificial Kiel Canal. It is, more or less, stagnant and salinity changes are weak. The natural stress of an estuary due to strong salinity changes is lacking and the number of brackish water species living there is high: from a survey in 1996, 38 macrozoobenthos species out of a total of 93 species have been classified as brackish species. It

has been found that the canal is a habitat for several endangered species such as *Streblospio benedicti*, *Gammarus duebeni*, *Heterotanaid oerstedii*, *Cyathura carinata* and *Embletonia pallida* (Leuchs *et al.*, 1999).

In the Dutch Wadden Sea, the restoration of salinity gradients in small sites has started. The management of the fresh water discharge in the Afsluitdijk has been adjusted to improve fish-migration into the IJsselmeer. However, knowledge on brackish-water habitats other than estuarine is scarce. Thus, in The Netherlands, a research program is being carried out on the possibilities of restoration of all kinds of aspects of estuarine (Ministerie van Verkeer and Waterstraat, 1996c).

5.15.9 Brief evaluation of the ecological situation of estuaries

Attempts to come to a qualitative/semi-quantitative evaluation of the ecological status of the estuaries under consideration have been made by Schuchardt *et al.* (1993b, Eider, Elbe, Weser, Ems) and Höpner (1995, Elbe, Weser, Ems).

The attempt by Höpner (1995) included nine features or functions, presenting a wider spectrum of estuarine characteristics. However, they have not been quantified and the evaluation was primarily relative between the estuaries, taking, also, the still existing nature protection potential into account. The results are shown in Table 5.12.

Schuchardt *et al.* (1993b) have compared the status quo (about 1985) with a historical reference situation (about 1880) for the indicators morphology/cross-sections, loss of foreland, deficit of dissolved oxygen and heavy metal concentration within the sediments which are partly quantifiable. The results are shown in Table 5.13.

Compared to the Wadden Sea, the estuaries are heavily stressed both due to the input of chemicals and due to morphological deterioration. However, there exist considerable differences between the degree of deterioration of the estuaries under consideration.

Some years ago, it was obvious that the Ems estuary has received less human impact than the other estuaries (except the Varde Å). The Eider estuary has been altered very strongly by morphological and hydrological changes due to storm flood protection and agricultural use, whereas the water quality is better than in the other estuaries. The Elbe estuary has been changed both due to deepening for sea-going vessels and as a result of coastal protection. The water quality is worse than in the other estuaries. However, natural conservation potential is still very high. In the Weser

estuary, deepening and canalization for shipping purposes is most significant and has altered the estuary greatly.

Taking the actual situation into account, this ranking must be changed considerably. The ecological status of the inner Ems estuary must be classified as heavily stressed regarding water quality, sediment budget, morphology and aquatic fauna, whereas the water quality parameters in the Elbe and Weser estuaries have been improved.

Data for the Varde Å are hardly available; thus, it cannot be evaluated in detail in the present paper. However, the Varde Å can be described as mostly natural in terms of hydrographic and morphological structure, although the marshes have also been strongly altered due to agricultural purposes.

Since some improvements of water quality have

tendencies existing, a further deterioration due to further deepening of the inner and outer estuaries and active habitat restoration (Gaumert, 1995; Schuchardt, 1997).

We will give a short summary of the most important or advanced schemes:

Ems estuary: irregular dredging of the inner estuary (Papenburg to Pogum) from 6.8 m to 7.3 m below high water for the transfer of ships from the Papenburg ship yard to Emden; extension of the Emden port; building of a storm-surge-barrier.

Weser estuary: Weser Tunnel at Dedesdorf; industrial district on the Luneplate; extension of Nordenham and Brake harbors; extension of the container terminal in Bremerhaven (CT IIIa and IV); filling of harbor basins in Bremen; canalizing and dredging of the river Hunte (km 0 - 21); deep-

biotope-diversity	Elbe	>	Ems	>	Weser
water quality	Ems	>	Weser	>	Elbe
sediment quality	Ems	>	Weser	>	Elbe
side-arms	Elbe	>	Ems	>	Weser
filter potential	Elbe	>	Ems	>	Weser
importance for avifauna	Ems	>	Elbe	>	Weser
importance for fish	?		?		?
importance for nature protection	Ems	>	Elbe	>	Weser
quality for recreation	Elbe	>	Ems	>	Weser

Table 5.12. Evaluation of the ecological status of the estuaries relative to each other (Höpner, 1995). (Höpner had also included the Oder river mouth). The assessment of the status of the Ems should be updated for recent changes.

been achieved during the last years (especially in the Elbe estuary), the deterioration of the physical environment (morphology, hydrology) can now be regarded as the most severe impact on the ecological situation of all four German estuaries.

5.15.10 Current and future activities within the estuaries

Looking into the future, further improvement of water quality, especially in the Elbe estuary, can be expected. The situation in the Ems estuary is unclear. Concerning morphology, there are two

ening of the outer Weser (Bremerhaven to the open sea) from 12 to 14.7 m below sea chart zero;

Elbe estuary: deepening of the lower and outer Elbe (Hamburg - Cuxhaven) from 13.5 to 16 m below sea chart zero; raising of dikes and storm surge protection constructions, shortening of dike length; filling of harbor basins in Hamburg; transformation of marshes into dumpsites and harbor areas (Altenwerder).

All these activities will perpetuate and increase the adverse effects on the ecological system 'estuary'.

	Eider	Elbe	Weser	Ems
morphology / cross-sections	very strong	strong	very strong	moderate (*)
loss of foreland	no	very strong	strong	moderate
deficits of dissolved oxygen	small	very strong	strong	moderate (*)
heavy metals in sediments	moderate	very strong	very strong	strong

Table 5.13. Evaluation of the ecological status of the estuaries (about 1985) related to a historical reference situation (about 1880) (Schuchardt et al., 1993b). (*): must, on the basis of recent developments, be categorized as very strong.

5.15.11 Recommendations

- Existing ecological targets for estuaries (e.g. Leeuwarden) must be detailed, taking into account the individuality of each estuary.
- Monitoring of ecological long-term changes, other than water quality and macrozoobenthos in the estuaries, is necessary.
- The tidal freshwater reaches should be integrated into the Leeuwarden definition of an estuary.
- Active restoration of estuarine habitats (especially shallow areas and foreland) is necessary in all estuaries under consideration. Problems linked to the artificial increase of the tidal range have to be given special attention.
- Consequences of further impact due to further deepening, barriers and harbor extension should be evaluated very carefully, taking into account the historical deterioration of the estuaries and the uniqueness of each estuary.
- Further improvement of water quality is necessary, especially, for the Elbe and Ems estuaries.
- In contrast to the other estuaries, the ecological status of the Ems estuary has strongly deteriorated in recent years. An improvement of both, water quality and morphology, is necessary.
- Active restoration of smooth gradients of salinity and tidal amplitude in small creeks along the Wadden Sea coast and the estuaries to improve brackish water habitats is important.

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